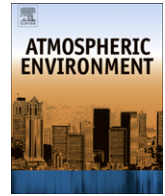


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# Atmospheric Environment

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## New Directions: Ash dieback and British carbon stocks<sup>☆</sup>



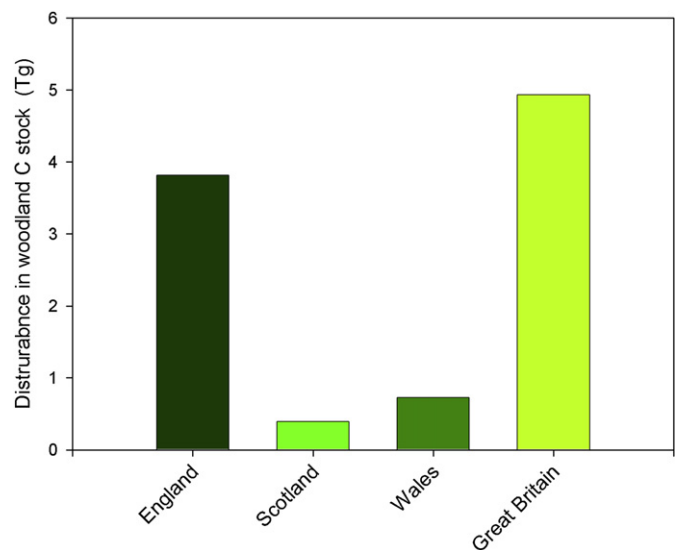
Confirmation of ash dieback caused by *Hymenoscyphus pseudoalbidus* (the sexual stage of the fungus *Chalara fraxinea*) in Great Britain in 2012 (Cromier, 2012) was a devastating blow for an arboreal stalwart of the British countryside and the many species that depend upon it. The full repercussions of widespread ash dieback for our woodland ecosystems remain to be seen, but the sheer number of ash trees (*Fraxinus excelsior*) across Britain means that a high infection rate (90%) (EPPO, 2010; Bakys, 2013) coupled with the mortality rate already reported in Lithuania (60%) (Vasaitis and Lygis, 2008) would represent a major disturbance to national carbon stocks.

The Forestry Commission estimate that there are some 126 million ash trees in Great Britain's woodlands, the bulk of these being in England. Together, they comprise 11% of total stocked area and have an estimated wood volume of more than 33 million cubic metres (Forestry Commission, 2012a). If 90% of British ash trees become infected, a 60% mortality rate would result in a disturbance to British woodland carbon stocks of around 5 million tonnes C (Fig 1). This is equivalent to >4% of the total vegetation C stock in Great Britain, or the entire vegetation C stock of Northern Ireland (Ostle et al., 2009).

Wind-blown spores allow natural spread of the disease by between 20 and 30 km per year, but more rapid spread may occur through transport of infected plants and wood products (Forestry Commission, 2012b). As of March 4th 2013, there were 391 confirmed infection sites in the UK, the bulk being in eastern England but with numerous infection sites also reported in Scotland, Wales and Northern Ireland (Forestry Commission, 2013). Infected trees under 10 years of age are likely to die within 2–10 years, with trees under 40 years old that are also infected with honey fungus dying within 3–5 years. As yet, there are insufficient data on ash dieback-induced death in trees more than 40 years old (Forestry Commission, 2012b).

Clearly, any actual net change will depend upon the rate at which ash cover is replaced by other vegetation and the form this regrowth takes (Pautasso et al., 2013). For woodland carbon stocks as a whole, the pathway of deadwood C cycling will be crucial in determining this net effect – increased input to litter and soil carbon stocks can be expected, while in some areas an increase in wood extraction for fuel use may occur.

As already called for in North America (Hicke et al., 2012), more research is urgently required to assess the potential impacts of such diseases on carbon stocks and net carbon fluxes in Britain and its



**Fig. 1.** Projected magnitude of carbon stock disturbance (Tg C) caused by ash dieback in Great Britain and its regions. Assumes 90% infection rate (Bakys, 2013) and 60% mortality rate (Vasaitis and Lygis, 2008). Values estimated using an ash wood density of  $0.6 \text{ g cm}^{-3}$  and a carbon content of  $0.45 \text{ gC g}^{-1}$  (Le Goff et al., 2004).

regions. If impact projections for major biotic disturbances, such as ash dieback, can be improved, then future policy and woodland management responses will be all the more effective.

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<sup>☆</sup> Something to say? Comments on this article, or suggestions for other topics, are welcome. Please contact: [new.directions@uea.ac.uk](mailto:new.directions@uea.ac.uk), or go to [atmos\\_env@uea.ac.uk](mailto:atmos_env@uea.ac.uk). [http://www.elsevier.com/wps/find/journaldescription.cws\\_home/246/authorinstructions](http://www.elsevier.com/wps/find/journaldescription.cws_home/246/authorinstructions) for further details.

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